Associations between air pollution and biomarkers of Alzheimer’s disease in cognitively unimpaired individuals

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Abstract

Background: Air quality contributes to incidence of Alzheimer’s disease (AD) although the underlying neurobiological mechanisms are unclear. This study was aimed to examine the association between air pollution and concentrations of cerebrospinal fluid (CSF) AD biomarkers and amyloid-β (Aβ) deposition.

Participants and methods

The sample included 156 cognitively unimpaired adults aged 57 years (61 at biomarkers assessment) with increased risk of AD from the ALFA Study. We examined CSF levels of Aβ42, Aβ40, p-Tau, t-Tau, neurofilament light (NFL) and cerebral amyloid load (Centiloid). A Land Use Regression model from 2009 was used to estimate residential exposure to air pollutants including nitrogen dioxide (NO2) and particulate matter (PM2.5, PM2.5 abs).

Abbreviations: AD, Alzheimer’s dementia; Aβ, Amyloid-β; APOE-ε4, APOLIPROTEIN E (APOE) epsilon 4 (+) allele; β, βeta coefficient; ALFA, ALzheimer and FAmilies; CI, Confidence Interval; CSF, Cerebrospinal Fluid; CL, centiloid values; LUR, land-use regression models; MRI, Magnetic Resonance Imaging; NFL, neurofilament light; NO2, Nitrogen dioxide; NOx, Nitrogen Oxides; PET, positron emission tomography; p-tau, phosphorylated tau; PM, Particulate Matter; t-tau, total tau; SD, Standard Deviation; SE, Standard Error.
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1. Introduction

Alzheimer’s disease (AD) is a slowly progressive neurodegenerative disorder clinically characterised by deterioration of episodic memory and successive impairment of additional cognitive domains, with behavioural changes impacting activities of daily living. The pathological hallmarks of AD include the extracellular accumulation of amyloid-β (Aβ) and intracellular aggregates of hyperphosphorylated tau (Frisoni et al., 2017). Since the number of individuals affected by AD is expected to rise due to the aging population, and there is still no cure for this disease, there is increasing interest in identifying modifiable environmental risk factors (Crous-Bou et al., 2017).

Air pollution is an emerging environmental risk factor for AD but knowledge on the effect of this exposure on the central nervous system is limited. Air pollution exposure estimated using land-use regression models (LUR) has been associated with cognitive decline (Tzivian et al., 2016). In the light of these findings, a few epidemiological studies have considered the role of air pollution exposure including nitrogen oxides (NOₓ) and particulate matter (PM) with respect to AD (Molinuevo et al., 2016). The ALFA cohort was established as a research platform aimed to characterize preclinical AD among 2743 cognitively unimpaired individuals, aged between 45 and 75 years, and enriched for family history of AD. Participants were recruited between 2013 and 2014. Impaired cognition was determined by cognitive performance falling outside the established cutoffs of the following cognitive tests: Mini-Mental State Examination (Blesa et al., 2001; Folstein et al., 1975) (MMSE), 26, or Memory Impairment Screen (Böhm et al., 2005; Buschke et al., 1999), 6, or Time-Orientatio subtest of the Barcelona Test II (Quinones-Ubeda, 2009) (TO-BTH), 68, or semantic fluency (Pena-Casanova et al., 2009; Ramier & Hecaen, 1970) (animals; SF), 12. The nested ALFA + study consists of a subset of participants of the ALFA parent cohort, who were invited to join based on their specific AD risk profile including AD parental history, APOE status and cognitive performance (Molinuevo et al., 2016). In the present study, we included the first consecutive 450 participants of ALFA + that were assessed between 2016 and 2019. From these, we selected only those individuals residing in the city of Barcelona in the same residence for at least the previous 3 years, and with reliable geocoded data leaving a final sample of 156 individuals. A flow chart depicting the final sample of the study is presented in Supplementary Fig. 1 (Fig S1).

The ALFA + study (ALFA-PFM-0311) was approved by the Independent Ethics Committee “Parc de Salut Mar,” Barcelona, and registered at Clinicaltrials.gov (Identifier: NCT02485730). All participants signed the study’s informed consent form that had also been approved by the Independent Ethics Committee “Parc de Salut Mar,” Barcelona.

2. Methods

2.1. Study population

Participants were selected from the ALFA + study, a nested longitudinal long-term study of the ALFA (Alzheimer and Families) cohort (Molinuevo et al., 2016). The ALFA cohort was established as a research platform aimed to characterize preclinical AD among 2743 cognitively unimpaired individuals, aged between 45 and 75 years, and enriched for family history of AD. Participants were recruited between 2013 and 2014. Impaired cognition was determined by cognitive performance falling outside the established cutoffs of the following cognitive tests: Mini-Mental State Examination (Blesa et al., 2001; Folstein et al., 1975) (MMSE), 26, or Memory Impairment Screen (Böhm et al., 2005; Buschke et al., 1999), 6, or Time-Orientatio subtest of the Barcelona Test II (Quinones-Ubeda, 2009) (TO-BTH), 68, or semantic fluency (Pena-Casanova et al., 2009; Ramier & Hecaen, 1970) (animals; SF), 12. The nested ALFA + study consists of a subset of participants of the ALFA parent cohort, who were invited to join based on their specific AD risk profile including AD parental history, APOE status and cognitive performance (Molinuevo et al., 2016). In the present study, we included the first consecutive 450 participants of ALFA + that were assessed between 2016 and 2019. From these, we selected only those individuals residing in the city of Barcelona in the same residence for at least the previous 3 years, and with reliable geocoded data leaving a final sample of 156 individuals. A flow chart depicting the final sample of the study is presented in Supplementary Fig. 1 (Fig S1).

To date, only one study has investigated in vivo cerebrospinal fluid (CSF) biomarkers related to neurodegenerative disorders among urbanite children and clean air controls (Lilian Calderón-Garcidueñas et al., 2015; Schikowski et al., 2015). Studies in post-mortem brain tissue provide biological support for the relationship between air pollution and AD. In autopsy samples from children and young adults who resided in high polluted areas, extracellular deposition of Aβ and phosphorylated tau (p-tau) were observed, which are considered hallmarks of AD (L. Calderón-Garcidueñas et al., 2015; Schikowski et al., 2020). Interestingly, no evidence of AD neuropathological change was found among non-exposed individuals (Lilian Calderón-Garcidueñas et al., 2012). To date, only one study has investigated in vivo cerebrospinal fluid (CSF) biomarkers related to neurodegenerative disorders among urbanite children and clean air controls (Lilian Calderón-Garcidueñas et al., 2016). Thus, the extent to which different air pollutants can alter in vivo AD biomarkers such as Aβ and tau levels in humans, is yet to be established.

Furthermore, when examining potential associations between air pollution exposure and AD biomarkers, it is important to consider the role of APOE-ε4 and the progression of the disease.

In the present study, we aimed to examine the association between air pollutants and CSF and positron emission tomography (PET) biomarkers related to brain Aβ deposition, tau pathology and neurodegeneration in cognitively unimpaired adult individuals at increased risk for developing AD. In addition, effect modification by APOE-ε4 and Aβ status in these associations was also investigated.

Participants have resided in the same residence for at least the previous 3 years. Multiple linear regression models were used to estimate associations between air pollutants and biomarkers. The effect modification by CSF Aβ status and APOE-ε4 carrier status was also assessed.

Results: A consistent pattern of results indicated that greater exposure to NOₓ and PM₂.₅ absorbance was associated with higher levels of brain Aβ deposition, while greater exposure to PM₁₀ and PM₂.₅ was associated with higher levels of CSF NIL. Most associations were driven by individuals that were Aβ-positive. Although APOE-ε4 status did not significantly modify these associations, the effect of air pollutants exposure on CSF NIL levels was stronger in APOE-ε4 carriers.

Conclusion: In a population of cognitively unimpaired adults with increased risk of AD, long-term exposure to air pollution was associated with higher levels in biomarkers of AD pathology. While further research is granted to elucidate the mechanisms involved in such associations, our results reiterate the role of air pollution as an environmental risk factor for AD.

Participants were selected from the ALFA + study, a nested longitudinal long-term study of the ALFA (Alzheimer and Families) cohort (Molinuevo et al., 2016). The ALFA cohort was established as a research platform aimed to characterize preclinical AD among 2743 cognitively unimpaired individuals, aged between 45 and 75 years, and enriched for family history of AD. Participants were recruited between 2013 and 2014. Impaired cognition was determined by cognitive performance falling outside the established cutoffs of the following cognitive tests: Mini-Mental State Examination (Blesa et al., 2001; Folstein et al., 1975) (MMSE), 26, or Memory Impairment Screen (Böhm et al., 2005; Buschke et al., 1999), 6, or Time-Orientatio subtest of the Barcelona Test II (Quinones-Ubeda, 2009) (TO-BTH), 68, or semantic fluency (Pena-Casanova et al., 2009; Ramier & Hecaen, 1970) (animals; SF), 12. The nested ALFA + study consists of a subset of participants of the ALFA parent cohort, who were invited to join based on their specific AD risk profile including AD parental history, APOE status and cognitive performance (Molinuevo et al., 2016). In the present study, we included the first consecutive 450 participants of ALFA + that were assessed between 2016 and 2019. From these, we selected only those individuals residing in the city of Barcelona in the same residence for at least the previous 3 years, and with reliable geocoded data leaving a final sample of 156 individuals. A flow chart depicting the final sample of the study is presented in Supplementary Fig. 1 (Fig S1).

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2.2. Exposure

Air pollution measurements were collected within the air pollution monitoring campaign of the European Study of Cohorts for Air Pollution Effects (ESCAPE; http://www.escapeproject.eu/) following established protocols (BeeLEN et al., 2013; EFfENTS et al., 2012). Levels of air pollution exposure including nitrogen oxides (NOₓ and NO₂) and particulate matter (PM₂.₅ [particulate matter with aerodynamic diameter <2.5 µm], PM₁₀ [<10 µm], PM coarse [PM₂.₅-10, i.e., coarse particulate matter, between 2.5 µm and 10 µm], and PM₂.₅ absorbance [PM₂.₅ light absorption, which constitutes a marker of black or elemental carbon (Cyrus et al., 2009)]) were measured in three different seasons (warm, cold, and one intermediate temperature season) to model an annual average for 2009. Given the high correlation between NO₂ and NOₓ (r = 0.90), we present results only for NO₂. Exposures containing fine particles (PM₂.₅, and PM₂.₅ absorbance) are particularly relevant given their ability to cross epithelial, endothelial, and neural barriers, producing direct and indirect brain damage (Block & Calderón-Garcidueñas, 2009). LUR
models were used to estimate individual levels of exposure to these air pollutants at the participant’s residential addresses reported in 2013–2014 (this information was not available for 2009). The model for 2009 was considered as a surrogate of long-term exposure until the time of data collection (2013–2014) (Vert et al., 2017). The LUR model estimated in 2009 is considered representative of the posterior years because, even though air pollution levels have slightly decreased in Barcelona over the years (from year 2009 to the time of enrollment in the study in 2013–2014), spatial differences across the city have remained similar (i.e. subject’s exposure globally decreased in the same proportion) (Vert et al., 2017). LUR models were based on geographical information system (GIS) and statistical methods.

2.3. CSF biomarker measurements

We measured CSF levels of Aβ42, Aβ40, phosphorylated tau (p-tau), total tau (t-tau) and neurofilament light (NfL). The CSF Aβ42/40 ratio was used as a biomarker of Aβ pathology, and the cutoff for Aβ positivity (CSF Aβ42/40 < 0.071) was determined by a Gaussian mixture model as previously described (Milà-Alomà et al., 2020). The collection of CSF biomarkers in ALFA + is described elsewhere (Milà-Alomà et al., 2019; Milà-Alomà et al., 2020). In brief, CSF t-tau and p-tau were measured using the electrochemiluminescence immunoassays Elecsys® on a fully automated cobas e 601 instrument (Roche Diagnostics, Rotkreuz, Switzerland). CSF Aβ42, Aβ40 and NfL were measured with the prototype NeuroToolKit on a cobas e 411 or e 601 instrument (Roche Diagnostics, Rotkreuz, Switzerland). All measurements were performed at the Clinical Neurochemistry Laboratory, Sahlgrenska University Hospital, Mölndal, Sweden. All measures of CSF AD biomarkers were treated as continuous variables where higher levels indicate more pathological levels except for the CSF Aβ42/40 ratio, where lower values indicate more pathology.

2.4. Positron emission tomography (PET) scanning

Levels of brain Aβ deposition were assessed by PET. A T1-weighted MRI and an [18F]flutemetamol PET scan were acquired in all participants (mean time difference 97.1 days; range [14–343]). The T1-weighted 3D-TFE sequence was acquired in a Philips 3 T Ingenia CX scanner with a voxel size of 0.75 × 0.75 × 0.75 mm3, FOV 240 × 240 × 180 mm3, sagittal acquisition, flip angle 8°, TR = 9.9 ms, TE = 4.6 ms, TI = 900 ms. PET imaging was conducted in a Siemens Biograph mCT, following a cranial CT scan for attenuation correction. Participants were injected with 185 MBq (range 166.5–203.5 MBq) of [18F]flutemetamol, and 4 frames of 5 min each were acquired 90 min post-injection. Images were reconstructed with an OSEM3D algorithm using 8 iterations and 21 subsets and with point spread function (PSF) and time of flight (TOF) corrections into a matrix size of 1.02 × 1.02 × 2.03 mm.

All PET images were pre-processed using a validated Centiloid pipeline (Klunk et al., 2015) using SPMM12 (https://www.fil.ion.ucl.ac.uk/spm/software/spm12/). PET frames were realigned, averaged and co-registered to their corresponding MRI scans. MRIs were then normalized to the MNI space together with PET images. We calculated the standardized uptake value ratio (SUVr) in MNI space using the target region provided in the GAAIN website (http://www.gain.org/centiloid-project) and the whole cerebellum as reference region. SUVr values were then transformed to the Centiloid scale as detailed elsewhere (Salvador et al., 2019). Centiloid values (CL) as a measure of brain Aβ deposition were treated as continuous variable where higher levels indicate higher Aβ burden.

2.5. Statistical analyses

Multiple linear regression analyses were conducted to examine whether levels of air pollutants (i.e. NO2, PM2.5, PM10 and PM2.5 absorbance) were associated with AD biomarkers (i.e. Aβ42/40 ratio, p-tau, t-tau, NfL and CL as a measure of brain Aβ deposition). In these models, biomarkers were assigned as dependent variables and air pollutants as independent variables. Each pollutant was tested for association with each biomarker individually. We excluded the extreme values of the outcome variables (i.e. biomarkers) defined as either those values that fell outside of 3 times the interquartile range below the first quartile (Q1) or 3 times the interquartile range above the third quartile (Q3). Levels of p-tau, t-tau and NfL were log10-transformed for normality. All models were adjusted by sex, age at recruitment and APOE-ε4 status. Age at recruitment was strongly correlated with age at biomarker assessment (r = 0.98, p = 2.2-e16). Standardized regression coefficients are reported through results (β). Standardized coefficients represent the mean change in the outcome (measurement unit is pg/mL) given a one standard deviation change in the exposure (measurement unit is μg/m3).

In secondary analyses, we examined the potential effect modification by CSF Aβ status and APOE-ε4 carriership. We conducted sensitivity analysis further adjusting by lifestyle and clinical risk factors for AD including years of education, physical activity, body mass index (BMI), smoking habits and family history of AD.

Statistical significance was set at P < 0.05. All analyses were conducted using R statistical software package, version 3.6.2 (http://www.r-project.org/).

![Fig. 1. Dot-and-whisker plot showing associations between exposure to air pollutants and levels of Alzheimer’s disease (AD) biomarkers. All estimates were adjusted by age, age and APOE-ε4 status. Standardized coefficients and 95% confidence intervals are reported. Note: NfL, neurofilament light; CL, centiloid values.](image-url)
3. Results

The main characteristics of the study population are shown in Table 1. Mean age of participants was 56.7 years at recruitment (SD = 5.2), 60.6 (SD = 4.9) at CSF extraction and 60.9 (SD = 4.9) at PET scan. Sex was equally distributed (55.8% were female), 48.7% were carriers of the APOE-ε4 allele, and 60.9% had a relative with dementia. Socio-demographic characteristics of the study population according to Aβ and APOE-ε4 status are described in Supplementary Table 1 (Table S1) and Table S2, respectively.

3.1. Association between air pollutants and AD biomarkers

Associations between exposure to air pollutants and AD biomarkers are shown in Fig. 1 and Table S3. Higher exposure to NO₂ was associated with higher CL (NO₂: β = 0.21, 95% CI = 0.05–0.37, P = 0.012). Positive associations were observed between exposure to PM₂.₅ and PM₁₀, and CSF NfL levels (PM₂.₅: β = 0.17, 95% CI = 0.02–0.31, P = 0.027; PM₁₀: β = 0.15, 95% CI = 0.01–0.30, P = 0.047). Additionally, exposure to PM₂.₅ absorbance was associated with centiloid scale (β = 0.16, 95% CI = 0–2.32, P = 0.050).

The direction of effects was consistently positive for all biomarkers except for the CSF Aβ42/40 ratio suggesting that air pollution was related to more pathological levels of all AD biomarkers analysed. (Fig. 1).

In sensitivity analyses further adjusting by lifestyle and risk factors of AD (i.e. family history of AD, years of education, BMI, smoking habits, and physical activity), the observed associations between air pollutants and AD biomarkers were slightly attenuated (Table S4). The associations between NO₂, PM₂.₅ absorbance and CL remained significant. Although the associations between PM₂.₅ and PM₁₀ and CSF NfL were no longer significant, we observed the same direction of effects and estimates of similar magnitude.

3.2. Effect modification Aβ status

Associations between exposure to air pollutants and AD biomarkers stratified by the CSF Aβ42/40 ratio (Aβ status) are shown in Fig. 2. Forty-five individuals were Aβ positive and 102 were Aβ negative. Aβ status modified associations between air pollutants (NO₂, PM₂.₅, PM₁₀ and PM₂.₅ absorbance) and AD biomarkers (CSF p-tau, t-tau and NfL) (P for interaction < 0.05). Only among individuals with a positive Aβ status, exposure to NO₂ was positively associated with CSF t-tau (β = 0.34, 95% CI = 0.04–0.64, P = 0.027), CSF NfL (β = 0.36, 95% CI = 0.12–0.60, P = 0.005) and CL (β = 0.36, 95% CI = 0.08–0.64, P = 0.012). Similarly, associations between PM₂.₅, PM₁₀ and PM₂.₅ absorbance and CSF t-tau and CL were consistently positive and limited to Aβ-positive participants (Fig. 2). When further adjusting by APOE-ε4 status, associations between NO₂, PM₂.₅, PM₁₀ and PM₂.₅ absorbance and CL remained positive and stronger among CSF Aβ-positive individuals to negatives (P-interaction < 0.05). Although we did not observe other significant interactions, positive associations between NO₂, PM₂.₅, PM₁₀ and PM₂.₅ absorbance and CSF t-tau and NfL were limited to the Aβ-positive group (Table S5).

3.3. Effect modification by APOE-ε4 status

Associations between exposure to air pollutants and AD biomarkers stratified by APOE-ε4 status are shown in Table 4. Seventy-six individuals were APOE-ε4 carriers and 80 were non-carriers. Associations were not significantly different between carriers and non-carriers of the ε4 allele of the APOE (P-interaction > 0.05). Stratified analysis showed that several air pollutants were associated with CSF NfL (NO₂: β = 0.25, 95% CI = 0.03–0.47, P = 0.025; PM₁₀: β = 0.25, 95% CI = 0.04–0.46, P = 0.023; PM₂.₅ absorbance: β = 0.24, 95% CI = 0.01–0.36, P = 0.042) and CSF t-tau (PM₂.₅ absorbance: β = 0.24, 95% CI = 0.01–0.48, P = 0.045) levels only among carriers (N = 76). In contrast, NO₂ (β = 0.26, 95% CI = 0.02–0.49, P = 0.034), PM₁₀ (β = 0.25, 95% CI = 0.03–0.47, P = 0.028) and PM₂.₅ absorbance (β = 0.31, 95% CI = 0.08–0.54, P = 0.009) showed associations with CL only among non-carriers (N = 80) (Fig. 3).

4. Discussion

In a population of late/middle-aged cognitively unimpaired adults with increased risk of AD, long-term air pollution exposure showed associations with biomarkers of AD pathology assessed in CSF and Aβ PET. The pattern of results was consistent suggesting detrimental effects of air pollution on AD risk. Specifically, greater exposure to NO₂ and PM₂.₅ absorbance was associated with higher levels of brain Aβ deposition (as assessed by Aβ PET), while PM₁₀ and PM₂.₅ exposure was positively associated with higher levels of CSF NfL, a marker of neuronal injury. Furthermore, in participants within the Alzheimer’s continuum, defined by decreased CSF Aβ42/40 ratio (positive Aβ status), greater exposure to NO₂, PM₁₀ and PM₂.₅ was related with greater tau-related biomarkers and NfL and NO₂ with tau pathology. Although APOE-ε4 status did not significantly modify these associations, the effect of air pollutants exposure on CSF NfL levels was stronger in APOE-ε4 carriers.
Additionally, the associations between exposure to air pollutants and CL were stronger among non-carriers.

To our knowledge, this is the first study analysing associations between estimated exposure to air pollutants and in vivo AD biomarkers assessed in CSF and Aβ PET in humans. These findings are biologically plausible given that evidence suggests that fine particles can reach the brain via the circulation, bypassing the blood–brain-barrier by direct translocation through the olfactory bulb (Oberdörster et al., 2004; Peters et al., 2006). Also, air pollution is among the most relevant sources of environmentally induced inflammation and oxidative stress, both involved in neurodegenerative processes (Block & Calderón-Garcidueñas, 2009; Brockmeyer & D’Angiulli, 2016).

Our results are in line with previous animal and human autopsy studies suggesting that air pollution may be adversely associated with characteristics of Aβ and tau-pathology (M Cacciottolo et al., 2017; Calderon-Garcidueñas et al., 2003; Lilian Calderón-Garcidueñas et al., 2002, 2012; Lilian Calderón-Garcidueñas, Herrera-Soto, et al., 2020; Park et al., 2020). Air pollution has also been linked to brain structural alterations of relevance for AD such as progressive atrophy of gray and white matter (de Prado Bert et al., 2018). A recent study found that increased neuroanatomic risk for AD (a structural brain MRI-based score reflecting high-dimensional grey matter atrophies in brain areas vulnerable to Alzheimer’s disease neuropathology) mediated the association between long-term exposure to ambient PM2.5 at residential locations and declines in episodic memory (Younan et al., 2020). In this context, a previous study within the ALFA cohort showed that air pollution was associated with global atrophy and reduced volume and thickness in AD vulnerable regions (Crous-Bou et al., 2020). Our findings may be underlying these associations and overall provide support for the link between air pollution and AD-related outcomes.

Interestingly, we observed that participants within the Alzheimer’s continuum (Sperling et al., 2013), hence already Aβ-positive, drove the associations with Aβ, tau and neurodegeneration biomarkers. AD has an extensive preclinical stage, which may be initiated 15 to 20 years before the emergence of clinical symptoms (Sperling et al., 2013). Aβ negativity, especially when defined with Aβ CSF biomarkers, may indicate that AD-pathology is not yet present (Palmqvist et al., 2016). These results suggest that the detrimental effects linked to air pollution occurred downstream to Aβ pathology. Air pollution will therefore contribute to the tau and neurodegenerative changes downstream to Aβ pathology, hence affecting only those that are already within the continuum. A possible explanation for this finding is that Aβ-positive individuals may be more vulnerable to air pollution effects. In this regard, it is worth mentioning that a high proportion of Aβ-positive individuals (76%) were also APOE-ε4 carriers which may be accounting for their Aβ status and contributing to a greater vulnerability to air pollution. In future follow-ups, it will be relevant to examine whether associations between air pollution and AD-biomarkers observed at this time-point.

Fig. 2. Association between exposure to air pollution and levels of Alzheimer’s disease (AD) biomarkers by amyloid status. Each point depicts the value of the AD biomarker of an individual and the solid lines indicate the regression line for each of the groups (positive (A+) and negative (A−) amyloid status. P-value for the interaction between air pollution and amyloid status is shown in the upper right corner of each plot. Results for stratified analysis are reported below each plot. All estimates were adjusted by sex, age and APOE-ε4 status. Standardized coefficients (β) are reported. Note: NfL, neurofilament light; CL, centroid values.
change as AD-pathology evolves in some individuals.

The lack of associations between air pollutants and brain Aβ deposition among carriers may be explained by the role of APOE in Aβ metabolism (Liu et al., 2013). APOE is related to the formation and aggregation of Aβ plaques (Kanekiyo et al., 2014), being Aβ deposition more prevalent among APOE-ε4 carriers compared to non-carriers (Reiman et al., 2009; Snellman et al., 2020). Thus, variability in biomarkers related to Aβ deposition could be mostly accounted for by APOE-ε4 allele while among non-carriers, other risk factors such as air pollution could be triggering or contributing to these pathological characteristics. This highlights the need to identify modifiable risk factors for AD. It is also relevant to consider the temporal progression of AD-biomarkers to interpret these findings. CSF Aβ alterations are amongst the first detectable pathophysiological events in AD while characteristics of tau pathology and neurodegeneration are thought to appear later (Clifford et al., 2016; Palmqvist et al., 2016). Exposure to air pollutants showed positive associations with CSF NfL and t-tau levels only in carriers, where pathology may be more advanced. Alternatively, carrying the ε4 allele may also contribute to a greater vulnerability to other risk factors, either genetic or environmental, which exacerbate or accelerate AD-pathology. Indeed, findings in mice showed that APOE-ε4 carriers had more PM-induced Aβ plaques compared to non-carriers (M Cacciottolo et al., 2017). Although we did not observe associations between air pollution and centiloid scale among APOE-ε4 carriers, associations with CSF t-tau and NfL were only present in this subgroup.

The current study must be interpreted in the context of its limitations. First, the sample was selected to include individuals at risk for AD, thus, although individuals were cognitively unimpaired, generalizations to the general population should be conducted with caution. For instance, ≈50% of the individuals were carriers of the APOE-ε4 allele while its frequency in the general population ranges between 10% and 20% (Singh et al., 2006). Nevertheless, analyses were adjusted by APOE-ε4 status. Second, air pollution exposure estimations were based on participants’ home address. Data on exposure to air pollution related to mobility or workplace is not available, thus, we cannot discard exposure misclassification. Third, air pollution exposure based on 2009 models was used as a surrogate of the period of data collection, which took place 4–10 years later. Nevertheless, as above mentioned, there is evidence indicating that the spatial distribution of air pollution in Barcelona has remained overall consistent in the past 20 years (Cesaroni et al., 2012; Wang et al., 2013). Fourth, due to the limited sample size of our study, rather than adjusting for multiple comparisons, we emphasized on the pattern and consistency of results.

Considering the consistent pattern of results observed, we believe that our findings reinforce the emerging role of air pollution as a major environmental risk factor for AD. The association between air pollutants and AD-biomarkers suggest that this exposure may contribute to the etiopathogenesis of AD. Future analyses in ALFA + will allow testing whether air pollution effects are related to the progression of the disease. Given the ubiquitous nature of this exposure, even if associated risk is of small effect size, reductions of population-level exposure would lead to decrease the global burden of the disease.

Author contributions

SA, MCB, JS and JLM were responsible for the initial plan, study design, conducting the study, and data interpretation. SA and MCB are responsible of statistical analysis, data interpretation and manuscript drafting. CM and KF were responsible for recruitment, data collection and data deprecation. MMA, MSC, GS, GK, HZ, KB were responsible for biomarkers data acquisition and preparation. NVT was responsible for genetic data acquisition and preparation. MC, MG and MN were responsible of exposure data estimation. All authors interpreted the results, and critically revised the paper. SA, MCB, JS and JLM are guarantors and had full access to all the data, including statistical reports and tables, and take full responsibility for the integrity of the data and the accuracy of the data analysis. The corresponding authors (MCB, JLM) attest that all listed authors meet authorship criteria and that no others meeting the criteria have been omitted.
Declaration of Competing Interest

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: [HZ has served at scientific advisory boards for Denali, Roche Diagnostics, Wave, Samumed, Siemens Healthineers, Pintelon Therapeutics and CogRx, has given lectures in symposia sponsored by Fujirebio, Alzceure and Biogen, and is a co-founder of Brain Biomarker Solutions in Gothenburg AB (BBS), which is a part of the GU Ventures Incubator Program (outside submitted work).

KB has served as a consultant, at advisory boards, or at data monitoring committees for Abcam, Axon, Biogen, JOMDD/Shimadzu. Julius Clinical, Lilly, MagQu, Novartis, Roche Diagnostics, and Siemens Healthineers, and is a co-founder of Brain Biomarker Solutions in Gothenburg AB (BBS), which is a part of the GU Ventures Incubator Program.

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GK is a full-time employee of Roche Diagnostics GmbH.

JLM is currently a full-time employee of Lundbeck and has before served as a consultant or at advisory boards for the following for-profit companies, or has given lectures in symposia sponsored by the following for-profit companies: Roche Diagnostics, Genentech, Novartis, Lundbeck, Orzyzon, Biogen, Lilly, Janssen, Green Valley, MSD, Eisai, Alector, BioCross, GE Healthcare, ProMIS Neurosciences, NovoNordisk, Zamviron, Cytos and Nutricia.

The rest of the authors have no conflict of interest to declare.

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Data statement

The data that support the findings of this study are available from the corresponding authors (MCB, JLM), upon reasonable request.

Appendix A. Supplementary material

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